Fire science for rainforests

Mark A. Cochrane*†

* Center for Global Change and Earth Observations, Michigan State University, 1405 S. Harrison Road, Room 101, East Lansing, Michigan 48823-5243, USA † Instituto do Homem e Meio Ambiente da Amazônia (IMAZON), Caixa Postal 1015, Belém-PA, CEP 66017-000, Brazil

Forest fires are growing in size and frequency across the tropics. Continually eroding fragmented forest edges, they are unintended ecological disturbances that transcend deforestation to degrade vast regions of standing forest, diminishing ecosystem services and the economic potential of these natural resources. Affecting the health of millions, net forest fire emissions may have released carbon equivalent to 41% of worldwide fossil fuel use in 1997–98. Episodically more severe during El Niño events, pan-tropical forest fires will increase as more damaged, less fire-resistant, forests cover the landscape. Here I discuss the current state of tropical fire science and make recommendations for advancement.

ires in the tropics generally bring to mind decades of media images of deforestation fires where slash-and-burn practices are used to convert lush rainforests into agricultural lands. However, beyond these intentional fires looms the burning of vast regions of standing rainforests as an unintended consequence of current land-use practices. Quantitative estimates of area burned are sporadic, incomplete or nonexistent for most tropical countries but are nevertheless shocking. Within southeast Asia and Latin America, out-of-control fires burned more than 20 million hectares in 1997–98 (refs 1, 2). This is equivalent to half of California and is far from a complete accounting of unintentional burning in the tropics, because burning in many regions has not been well quantified.

Some perspective is necessary to grasp the scope of the problem faced in the tropics. The fire season of 2000 was one of the worst on record in the United States, with nearly 3.4 million hectares burned and over 1.4 billion US dollars in fire suppression costs (see the National Interagency Fire Center website, http://www.nifc.gov/ stats/). By way of contrast, in 1997-98, Indonesia was faced with 8 million hectares of burning and a fire-fighting budget, including foreign aid, of only \$25 million (ref. 3). The story is similar across the tropics: 3 million hectares burned in Bolivia⁴, 2.5 million hectares burned throughout Central America², and in Brazil, 5 million hectares burned in a single Amazonian state⁵. The ubiquitous nature of fire in the tropics is evident in satellite imagery. Over the years, several satellites have been pressed into service as firedetection devices (Box 1). These and newer, more adept, satellites show a landscape that appears to burn every year (see http:// earthobservatory.nasa.gov/Observatory/showqt.php3).

Throughout the last century, great efforts and vast resources have been applied to understanding and managing fire in forests. However, in the tropics, demographic and land-use changes have only recently made fire a matter of serious concern⁶. Consequently, critical knowledge and resources have been lacking. Fire science has come late to the tropics.

With so much existing knowledge about fire science it is fair to ask whether tropical fires merit specific study. Fire itself is simply a mixture of heat, fuel and oxygen. The particular mixture of these three components determines whether and how things burn. Integration of fire with the physical and biological components of the Earth leads to a remarkable amount of complexity in fire behaviour and effects. Add to this the evolutionary, economic and social context under which fire operates and it becomes clear that while the chemistry of fire may be universal, its effects are not. This review attempts to explain the present understanding of fire in tropical rainforests and outline areas needing further development.

Impacts of fire

Ecosystem effects

Burning of tropical evergreen rainforests alters forest composition and structure^{7,8}. Common tree species suffer the greatest total mortality⁹ but rare species are most likely to be locally extirpated⁷. Even 15 years after burning, forests show no evidence of regaining lost species⁹. Prospects for species recovery are diminished because surface fires reduce seed availability by 85% in the litter layer and 60% in the upper 1.5 cm of soil¹⁰, while flowering and fruiting of trees in and near burned forests also decreases¹¹. Conditions heavily weight post-fire regeneration toward wind-borne, light-demanding pioneer species. Unburned forest islands^{7,10} and gallery forests¹² are key seed sources for post-fire recovery; however, recurrent fires rapidly reduce the size and density of surviving unburned forest fragments⁷ and kill regenerating vegetation, further depleting the prospects for recovery of mature forest species¹⁰. Grasses and small vines also invade burned forests, slowing regeneration and increasing forest flammability^{8,13–15}.

Animal responses to fire are poorly understood but are complex and dynamic, with some groups, including small mammals, reptiles and amphibians, increasing in abundance or species richness¹⁶ while other groups decline (for example, birds, insects)^{16,17}. Most animals initially escape the fires but die through starvation or predation if they cannot find shelter and establish new territories².

With millions of hectares of fire-damaged forests spanning the tropics, studies of natural and managed ecosystem recovery are needed. In particular, the rates at which these damaged rainforests recover their near-fire-immune status need to be determined. Potential intervention or management strategies to enhance ecosystem recovery processes should also be assessed.

Emissions

Net annual carbon emissions from tropical wildfires are extremely variable, ranging from 7.5 to 70 Mg ha⁻¹ depending on previous fire and land-use history¹³. Emissions from unplanned tropical peat fires are more severe, possibly exceeding 300 Mg Cha^{-1} (ref. 18). Few estimates of total atmospheric carbon emissions from tropical forest fires exist but include 4.6 Tg from Mexico¹⁹, 23.2 Tg from Roraima, Brazil²⁰, and 810–2,570 Tg from tropical peat fires in Indonesia¹⁸ for 1997–98 El Niño fires. These three fires alone may have equalled 41% of world fossil fuel emissions²¹. Millions more hectares are known to have burned in southeast Asia¹, Africa, Central and South America². The vast majority of unconsumed biomass from killed trees will also slowly release carbon through decay processes^{1,7}.

Smoke from tropical forest fires causes human suffering and death. Across the tropics, hundreds have lost their lives in smoke-related accidents, including ship, automobile and plane crashes^{1,2}.

review article

review article

Box 1

Fire detection from space

Fire is a chemical reaction that releases energy in the form of heat and light. Burning changes the structure of vegetation and converts fuels into a host of combustion products that can remain in place (such as char and ash) or be transported from the site (such as aerosols). Conditions during and after fire change the spectrum of reflected and emitted energy from an affected area. Earth-observing satellite sensors can potentially make use of these changes to detect fires, quantify areas burned and determine the composition and distribution of smoke plumes.

Ideally, fires should be mapped by low-cost, frequently collected satellite imagery, but determining which tropical forests are burning or have burned has not turned out to be a simple exercise. Fire detection becomes a challenge of balancing missed detections of real fires and erroneous detections of false fires⁸⁰ (such as sun glint, clouds, bare soil surfaces). High-resolution satellites including the Landsat Enhanced Thematic Mapper (ETM+) and Systeme Probatoire d'Observation de la Terre (SPOT) can be used to detect fires but their relatively high cost, and limited spatial and temporal sampling, limit their usefulness for active fire detection⁸⁰. Frequently used satellite systems include the Advanced Very High Resolution Radiometer (AVHRR), the Geostationary Operational Environmental Satellite (GOES) and the Defense Meteorological Satellite Program Operational Line Scanner (DMSP-OLS) satellites because of their frequent coverage of the Earth's surface. Although originally designed for cloud detection, each sensor has different strengths and limitations for fire detection⁸¹. Issues include sensor sensitivity and saturation characteristics, spatial resolution, number of observations per day and processing methods among others². Several algorithms have been used to estimate the area burned with varying results. Threshold and contextual classifiers tend to detect different fires⁸². Within contextual classifications, different methods lead to differing detection capabilities of different types and sizes of fires⁸⁰, which is important in the tropics because deforestation fires are very hot and may smoulder for as much as a week⁶⁸ while forest fires can be of low intensity and low temperature¹³ and fires in open vegetation types, such as cerrado and pastures, can burn intensely but cool quickly⁸³. Where low-intensity fires are widespread over a region, they can even influence discrimination statistics of contextual classifiers in such a way that prevents detection of normally detected intense fires⁸⁰.

Newer satellite sensors are expanding these capabilities in monitoring the extent and frequency of fires⁸⁴. The Tropical Rainfall Measuring Mission-Visible and Infrared Scanner (TRMM-VIRS), may be particularly useful for detecting unusually large or hot fires because of the reduced sensor saturation of its larger sampling area and potential for discriminating fires from other hot surfaces⁸⁵. The Bi-Spectral Infrared Detection (BIRD) mission is currently testing a new generation of infrared array sensors that provide surface temperature, area, and geo-location of fires with a spatial accuracy of 300 m, better analysis capabilities for High Temperature Events (HTE) and the ability to discriminate between smoke and water clouds^{86,87}. The higher saturation levels of Moderate Resolution Imaging Spectroradiometer (MODIS) satellite sensors provide better detection and monitoring capabilities for active fires than previous satellite platforms. In addition, improved spatial and radiometric resolutions of these satellites are useful for investigating aerosols in smoke plumes²⁷ and enhance the accuracy of burned area estimates⁸⁶. Both coarse⁸⁸ and fine⁸⁹ spatial resolution sensors are useful for mapping forest burning in the tropics. Burn signatures are transient in imagery owing to rapid forest regeneration⁹⁰ but may be detectable for up to two years with some techniques⁸⁹. ERS-2 Synthetic Aperture Radar (SAR) imagery is also promising for mapping burn scars in tropical evergreen forests because of its ability to penetrate clouds and smoke⁶⁴; however, detection capabilities can be confounded by moisture, making knowledge of site weather information crucial.

Thousands more died from smoke-related illnesses from forest fires in Indonesia²² and Brazil²³. Health effects depend on the concentration, constituents and length of exposure to smoke²⁴ but include respiratory and cardiovascular difficulties among other illnesses²². The complex mix of particles, liquids and gaseous compounds released depend upon the type and efficiency of burning. Fire emissions have been studied and quantified for deforestation and savanna fires225 but not for tropical forest fires. Emissions may change greatly between day and night because many of the fires then convert from flaming to smouldering combustion¹³. Emissions can be trapped beneath the rainforest canopy, potentially leading to toxic conditions. In addition, smoke plumes from biomass burning release large amounts of aerosols to the atmosphere, a substantial fraction of which is black carbon. In part owing to the black carbon fraction, the mix of aerosols in the tropics reduces the average solar radiation absorbed at the Earth surface by up to $35 \,\mathrm{W}\,\mathrm{m}^{-2}$ and may double the heating of the lower 3 km of the atmosphere²⁶. These changes affect atmospheric stability and cloud formation and can ultimately reduce rainfall over large areas²⁷⁻²⁹.

Development, deforestation and fire

Development of forested areas has historically been synonymous with deforestation. The advent of slashing and burning tropical forests for agricultural purposes was thousands of years ago³⁰. Periodic clearing and abandonment of tropical rainforests is potentially a sustainable practice, if rotation rates are long enough, but can only support low population densities. Higher population densities, extensive forest clearing and extreme drought conditions of periodic mega-El-Niño events may have resulted in widespread forest fires, frequent and severe enough to have influenced the distribution and migration of indigenous people³¹. In the modern era, tropical forests are the 'frontier' and, as always, fire is the tool of choice for clearing the land. Fire is very useful because it quickly and effectively reduces the biomass of newly cleared forests to nutrientrich ash that can fertilize crops. Hard-won agricultural lands are managed using fire to keep forest regrowth at bay, reduce pests, and promote forage growth for domestic animals. Fire is needed to create charcoal, reduce trash and debris piles, and for cooking food.

Fire becomes a problem, however, when it escapes from its intended purpose and causes unanticipated economic and environmental damage. Heavy fire use has costs as well as benefits. Escaped fires destroy crops, pasturage, timber, infrastructure (for example, fence posts, buildings and so on), and livestock, at times leading to severe economic hardship for local populations^{2,32}. Ill-timed fires can also spoil newly cleared lands by under- or overburning the slashed trees, resulting in reduced productivity or abandonment. Careless fire use and arson can lead to extensive fires in logged and unlogged forests and reduce the profitability or attractiveness of large-investment, high-profit permanent agriculture.

Fire and tropical forests

The crux of the fire problem in tropical rainforests is not so much the introduction of fire into these ecosystems but the frequency with which they are being burned. Historical records^{1,33,34} and charcoal in soil profiles^{35–38} show that tropical forest fires, even in the wetter forests, are not unprecedented. Fire can be considered endemic but rare in tropical rainforests¹⁴ with return intervals of hundreds if not thousands of years³⁹. Wetter forests burn less frequently but are more vulnerable to fire than drier forests because they have thinner protective layers of bark⁴⁰ and suffer much higher mortality rates from fires. Infrequent fire disturbance has left tropical rainforests evolutionarily ill-adapted to current patterns of burning^{40,41}.

Fire susceptibility in tropical forests occurs largely because of moisture stress, during periods of extensive drought, when normally moist fuels dry out and become potentially flammable. However, closed-canopy evergreen rainforests are remarkably resistant to drought³⁵. Even after months without rain, these forests can

maintain an evergreen canopy and high sub-canopy humidity levels, making sustained combustion impossible^{35,39}. This apparent fire-immunity results from the effective trapping of transpired moisture such that most of the ambient forest humidity is derived from the trees themselves⁴². Resilience to climatic stress through moisture recycling^{43,44}, enhanced by the deep-rooting capacity of many tree species⁴⁵, allowed many tropical forests to persist through severe droughts of previous glacial periods⁴⁶. But despite adaptations to drought conditions, rainforests can and do burn. Both natural and anthropogenic disturbances to forest canopies decrease the ability of forests to maintain moisture, making them more vulnerable to fire. In anthropogenic landscapes, fire is a continual presence, making it only a matter of time until a sufficiently intense drought opens up the forest to fire².

Fire occurrence in tropical forests is largely associated with forest edges^{47–51} and is not randomly located throughout. Although tropical regions have the highest density of lightning strikes⁵² and some associated tree mortality, these events are usually associated with heavy rainfall and rarely lead to forest fires⁵³. Other natural causes of fire such as volcanic eruptions and burning coal seams³⁷ may be regionally important but, as in most locations throughout the world, fire in the tropics is primarily associated with human activity^{34,54} and influence on land cover^{47–51}.

Land cover effects

Tropical land cover is rapidly changing⁵⁵. Road construction in recent decades has provided access for millions of settlers to previously remote and inaccessible forests⁵⁶. Deforestation has necessarily followed. Forest clearance fragments the remaining forests^{55,57}. Resultant forest edges are buffeted by winds and desiccating sunlight. These edge effects lead to structural changes, including increased mortality of trees, decreased living biomass⁵⁷

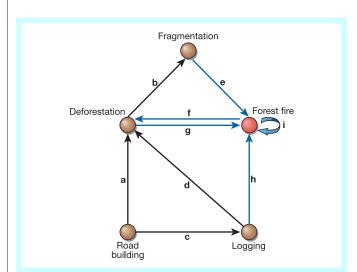


Figure 1 Diagram of interrelationships between tropical land-cover changes and forest fires. Arrows indicate forcing of each node upon others in the system. Blue arrows directly affect forest fire occurrence; black arrows indirectly influence forest fire occurrence. Events **a**–**i** are: **a**, Road building results in forest access that is strongly associated with deforestation⁵⁶. **b**, Deforestation fragments the remaining forests, increasing amounts of edge^{55,57}. **c**, Road building and paving directly affect transportation costs and area of economic accessibility⁷⁸. **d**, Logging results in limited amounts of deforestation for roads and log landings. Post-logging colonization can increase deforestation⁵⁹. **e**, Forest edges suffer biomass collapse and microclimate changes⁵⁷ making them susceptible to frequent fires⁴⁸. **f**, Repeated forest fires can lead to unintentional deforestation¹³. **g**, Deforestation and pasture/land maintenance fires result in many accidental forest fires can create a positive feedback cycle where recurrent fires become more likely and more severe with each occurrence¹³.

and increased fuel loads (that is, woody debris) 58 . These changes predispose these forests to fire 48,49 (Fig. 1).

Selective logging is another major landscape modifier in the tropics. The combination of human access provided by logging roads⁵⁹ and the forest damage caused by logging activities³⁹ make logged forests extremely vulnerable to fire⁶⁰. This vulnerability may last for decades after the logging activities have ceased⁶¹ (Box 2).

Given the juxtaposition of fire-prone forests and fire-dependent agricultural lands, forest fires are almost inevitable. In new frontier areas with relatively low quantities of cleared land, extensive undamaged forests make large forest fires nearly impossible. However, as fire-prone agricultural land development continues, the forest becomes increasingly fragmented and the whole landscape becomes conducive to fire propagation². The forest fires that do occur are often edge-related, moving into forests from deforested lands^{48,51}. These fires can significantly alter fire regimes kilometres from forest edges (Fig. 2). Fire frequency becomes a function of distance from deforested forest edges^{48,62} and fire severity increases with frequency¹³. In the absence of other modifying disturbances, these forests will continue to erode, with isolated fragments collapsing, unless future fires can be prevented⁶³. This is a long-standing process in the African tropics⁴³. Within the Brazilian Amazon alone, 26 million hectares of forest could currently be undergoing this process, leading to the eventual release of 3.9 Pg C (ref. 62).

Fire behaviour

Initial fires in relatively intact rainforests do not seem to be severe. Often progressing as slowly creeping ribbons of flame, 10 cm or so in height, burning little besides fallen leaf litter, these fires are still capable of killing 23–44% of the trees >10 cm in diameter at breast height (DBH)^{7,8,10,15,40,60}. Fire propagation in tropical forests is largely controlled by variations in ambient relative humidity

Box 2 Selective logging and fire

Millions of hectares of tropical forests have already been logged and millions more are being logged each year^{61,91,92}. Logging in tropical forests is often selective in that only a few valuable trees are extracted, leaving most of the forest in place. Impacts of selective logging vary with extraction intensity and management practices^{93–95}, but can be substantial. Selectively logged forests can recover to pre-harvest levels of biomass if left undisturbed. However, many forests are revisited several times when loggers return to harvest additional tree species as regional timber markets develop⁹³. These forests become very degraded and may have 40–50% of the canopy cover destroyed during these logging operations⁹³. Selective logging increases fire susceptibility^{99,60}, and can catalyse fire use and deforestation by opening up unoccupied and protected lands to colonization⁵⁹.

Fire severity in logged forests can be high^{2,8,61} owing to the large amount of available fuel in the form of slash piles and collateral damage caused by the logging operations^{39,93}. When weighted by area, logged forests near forest edges may not be more likely to burn than unlogged edge-forests but burn severity and fire spread rates can be higher, leading to deeper penetration and increased damage by fire (refs 34, 50, 96; see also http://www.fire.uni-freiburg.de/se_asia/ background/sea_5.html). In landscapes heavily affected by logging and fire, the relationship between fire frequency and distance from deforested edges can decrease or disappear completely because the entire forest will have similar fire-return intervals, often less than 10 years (ref. 48). Of deforestation, fire use or selective logging, none by itself is likely to create severe fire problems in tropical landscapes but the synergy and interaction between these land-uses when present together can promulgate fire throughout the landscape and rapidly degrade forests^{2,47-50}.

review article

(Box 3)^{13,39,64}. Although fire intensity is very low (4–55 kW m⁻¹), slow spread rates result in considerable heating of tree bark surfaces¹³. Most species in closed canopy evergreen forests have thin bark³⁹, so fires effectively girdle and kill them. Larger, thickerbarked trees and others without vulnerable cambium layers (for example, palms) survive^{13,15,39}. In some forest types with root mats or deep organic layers, ground fires can accompany surface fires. These fires may consume everything down to the mineral soil and cause near-complete mortality^{15,35}. Fires in logged or previously burned forests are more severe, owing to greater fuel loads and lower humidities^{7,40,60}, with fireline intensities of 82–728 kW m⁻¹ and more extreme fire behaviour¹³.

Although the chemistry and physics behind fire in tropical forests is no different from that in temperate or boreal forests, modelling fire susceptibility, behaviour, and effects in these forests will be more difficult owing to differences in tropical ecosystem characteristics. Closed-canopy tropical forests are species- and fuel-rich ecosystems that can regulate subcanopy humidity and hence the availability of fuels containing a variable mixture of numerous chemical compounds (for example, aliphatic and aromatic hydrocarbons, alcohols, aldehydes, gums, sugars, terpenes, fats, waxes and oils)⁵². This ecosystem modulation of atmospheric humidity modifies fire behaviour in complex ways by changing the ignitability, sustainability and combustability of fuels with varied structure and chemical composition.

Modelling of fire susceptibility in tropical evergreen forests is in its infancy. Fire susceptibility models must incorporate not only ambient weather conditions but also forest type, structure and disturbance history, as well as the availability of and ability to tap into deep soil moisture. Current conceptual models of susceptibility, the time since rain³⁹ and leaf-shedding^{65,66} fail to describe or provide the mechanistic understanding of forest moisture dynamics that are necessary to explain the behaviour of observed forest fires¹³. The deep-rooting nature⁴⁵ and tight hydrologic cycling⁴¹ of these forests make fire-susceptibility modelling challenging. Current models³² provide appropriate drought indices for these forests but cannot address the likelihood of actual fire ignition or spread owing to conceptual flaws that ignore the feedbacks of the forest upon the fire environment.

In the tropics, as elsewhere, models of fire spread and behaviour

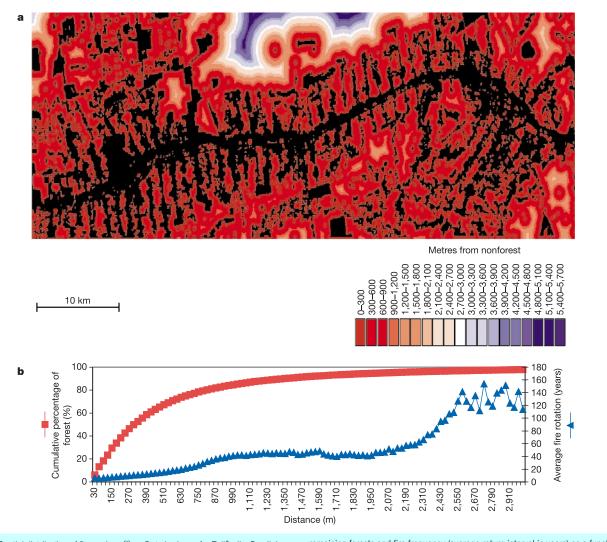


Figure 2 Spatial distribution of fire regimes⁶². **a**, Data is shown for Tailândia, Brazil, in 1997. Black areas are previously deforested, while other colours represent standing forests suffering different levels of fire impact. Forests in red are burning too frequently to persist as tropical evergreen forests and are turning into grassland and scrub¹³ in a continual erosion process^{43,63}. **b**, The graph shows both cumulative percentage of

remaining forests and fire frequency (average return interval in years) as a function of distance from deforested edges in the eastern Amazon. Tropical forests around the world are affected similarly but to different degrees depending on climate, fire load, and previous land cover changes or disturbances.

review article

will require appropriate fuel models. Such fuel models have been standardized for many ecosystems and forest types⁶⁷ but not for tropical forests. Fires in tropical forests frequently occur in surface fuels, so understanding the characteristics and compositions of litter systems will be essential for predicting burning characteristics. Litter arrangement, loading, moisture, type and composition are all important modifiers of fire behaviour. Although there are a few quantitative descriptions of fuel quantities and size distributions^{13,39,68}, there do not appear to be comparable studies of fuel composition and chemistry. Litter production and decay rates vary widely in the tropics. Furthermore, litter from different species have large variations in energy values and flammability. Therefore, the phytosociological mix of litter-generating species can influence the flammability, sustainability, spread and heat-release characteristics of fires⁵². The fuel structures of forests with deep root mats or peat layers must also be understood because locally devastating ground fires in these substrates can have globally significant impacts¹⁸. With their extremely species-rich complements and varieties of canopy, surface and subsurface fuel structures, tropical forests will probably require a whole new set of fuel models to parameterize existing fire models such as BEHAVE Plus (http://fire.org/cgi-bin/ nav.cgi?pages=behave&mode=8) and FARSITE⁶⁹. Dynamic relationships between fuels and forest moisture conditions will also be needed to address rapidly changing fuel availability and the potential involvement of green fuels.

Fire severity can be subjective but is often confused with intensity. This is understandable because severe fires in temperate and boreal forests are often intense surface or crown fires. The majority of tropical fires, however, are neither fast-spreading nor intense. They are severe nevertheless because the impact to the forest is quite large^{7,8,10,15,40,60}. A severe fire can be rapidly spreading if it is intense but can also be of low intensity if it is slow-spreading. Relative severity for an ecosystem will depend on its specific adaptations, fortuitous or evolved, to the fires that are experienced. Prescribed burning is neither effective nor useful in tropical rainforests. Although burning favours the establishment of some commercial timber species⁷⁰ and temporarily reduces growth-inhibiting vine loads⁷¹, collateral damage from fire use outweighs any benefits^{70,71}. Fires are not useful for managing fuels either, because even very-low-intensity fires cause high mortality, creating more fuels than they consume¹³. Specific fire-protection and suppression recommendations are needed that can provide measurable results in reducing fire frequency.

Fires in tropical rainforests range from easily extinguished litter fires to nearly impossible-to-extinguish ground fires. In recurrent wildfires¹³, average fire intensities are ten times greater (30 versus $307 \,\mathrm{kW \,m^{-1}}$) and can spread twice as fast (0.25 versus $0.52 \,\mathrm{m \,min^{-1}}$). These limited data should not be considered to limit the range of possible fire conditions in tropical forests. In particular, the potential for landscape-level interactions among multiple fires to create large-scale, rapidly spreading conflagrations must be considered. Mass fires, also known as 'firestorms', can spread rapidly and lead to fire-induced weather, including highvelocity winds, lightning, and extreme fire behaviour (for example, fire whirls)⁷². There is historical precedence⁷³ for widespread logging damage, previous fire damage and widespread clearing fires combining to create mass fires in other forests, with devastating environmental, economic and human consequences. Regions with

Box 3 Fuels and moisture

Fuels are commonly divided into different size classes that have been determined to have specific moisture flux characteristics, specifically 1-h, 10-h, 100-h and 1,000-h fuel classes (0–0.62, 0.62–2.54, 2.54–7.62 and >7.62 cm). The number reflects the amount of time a fuel particle of a given size requires to reach 63% of equilibrium after a change in ambient moisture conditions⁹⁷. Smaller fuels (1-h) can either dry or become moist relatively quickly while larger fuels (1,000-h) take a long time to dry or recover moisture. Fuels 'available' for a given fire are simply those substances that require less heat to reach the point of combustion than is being radiated by the surrounding environment. Moisture increases the amount of heat necessary to reach the point of ignition. Sustained combustion occurs when ignited fuels conduct sufficient energy into adjacent fuels to start combustion before the currently burning fuels are consumed. If there is a continuous layer of fuels where this process can occur, then the fire will spread.

Closed-canopy forests are protected from the wind, and fire intensities and spread rates in such forests are also kept low by the high moisture content of fuels. But even at low intensities, slowmoving fires are severe because of the mortality caused by the long fire-contact times and the resultant heat transfer at the base of trees¹³. The moisture that protects tropical forests from fire is therefore also one of the reasons for the high mortality rates experienced when these forests do dry out enough to catch fire.

Owing to standard fuel monitoring practices in temperate forests, 10-h fuel-sticks have been used to evaluate fuel moisture and fire susceptibility in tropical forests^{35,39,60}. However, in many tropical forests, 1-h fuels carry the fires while 100-h and 1,000-h fuels contain smouldering fires throughout moister periods (for example, nights, short rains and so on), allowing fires to persist and spread¹³. Consequently, moisture levels in the 10-h fuel class may actually be the least appropriate for characterizing fire susceptibility in tropical forests.

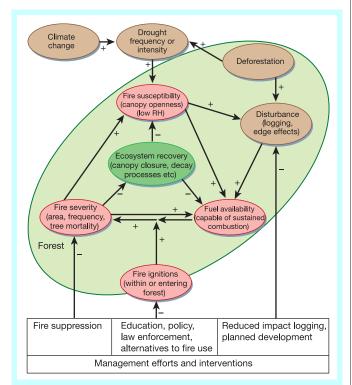


Figure 3 Positive and negative feedbacks controlling fire processes in tropical forests. Positioning indicates whether they occur within or outside the forest (green shading). Items in red control fire occurrence or behaviour. Items in brown modify the potential fire environment. Green indicates ecosystem processes acting in opposition to fire, specifically regrowth, canopy closure and decay of fuels. The management efforts and interventions box indicates how and where human actions can diminish tropical forest fires. Climate change encompasses effects of increased CO₂, land cover change, and aerosol loading that result in regional drying in the tropics^{26,27,79}. RH, relative humidity.

extensive logging- and fire-damaged forests that experience widespread and heavy fire use (such as East Kalimantan, Indonesia^{1,15,61}) are at greatest risk for such catastrophic fires.

Environmental impacts and climate feedbacks

Forest fires, deforestation and other forest disturbances effectively thin forests, reducing the amount of vegetation transpiring water. Reduced transpiration results in lowered local atmospheric humidity levels and increases the probability of future forest-fire occurrence⁷⁴. Decreased atmospheric moisture can negatively affect montane cloud forests by reducing the extent of and increasing the minimum altitude of downwind cloud fields⁷⁵. Regionally, transpiration from tropical forests is important for downwind precipitation, contributing 25% or more of annual rainfall in some regions^{43,44}.

Tropical forest fires reduce the ability of affected forests to retain water, exacerbating flooding, erosion and seasonal water shortages². Smoke-borne aerosols from fires disrupt normal hydrological processes and reduce rainfall^{28,29}, potentially contributing to regional drought. Through the radiative effects of smoke-derived aerosols, fires reduce the relative humidity and temperature gradients of the lower troposphere²⁷, disrupting normal cleansing processes and increasing average residence times of smoke particles in the atmosphere²⁶. Smoke plumes from tropical fires also increase the number and intensity of positive cloud-to-ground lightning strikes, long after and thousands of kilometres away from fires, altering 'natural' fire processes within and outside the tropics⁷⁶.

After a fire, leaves from newly killed trees quickly reblanket the ground while simultaneously opening the canopy, resulting in a more easily dried and fuel-laden forest^{7,13}. Whether burned^{7,14}, logged⁶⁰ or subjected to wind-throw and other edge effects^{48,49}, these forests become fire-susceptible after a few weeks without rain instead of several months⁷. Increased fire intensity and heat-transfer capacity of recurrent fires make them deadly to even larger thicker-barked trees¹³. Forest burning can create a positive feedback whereby more frequent and severe fires result in complete deforestation¹³ (Fig. 3).

Understanding fire in tropical rainforests

The existence of tropical forest fires has been periodically rediscovered and forgotten over the last century. Since Indonesia's large 1983 forest fires, studies of fire potential and effects have rapidly accrued in the literature but it is disturbing that Aubréville⁴³ was able to report much of this knowledge when he described fire effects in African forests in 1947. To move forward, tropical fire science needs to go beyond observation and description to provide increased understanding of the interactions between fire and tropical forests so that effective policy, management, modelling and monitoring efforts can be created.

Landscape-level knowledge of fire occurrence, behaviour and history is necessary for managing fire in the tropics. Tropical landscapes are a dynamic collection of intact, cleared, damaged and regenerating forests. To do more than chart the frequency and timing of burning at the landscape level, it will be necessary to relate fire-detection information to up-to-date land-cover maps. For ecological understanding of the importance of fire it will be necessary to know what is burning, how often and when. The need for high sampling frequency using high-resolution imagery currently makes the mapping of forest fires in the tropics expensive and labour-intensive². Where possible, such work should be combined with ongoing land use and land cover classifications to maximize the efficiency of resource use and the synergy of site information (for example, deforestation, selective logging, previous burning, forest regrowth and so on).

Fires have an impact on millions of hectares of rainforest², increase deforestation rates by as up to 50% in some regions¹³ and cause billions of dollars in damages in the tropics^{1,2}. To avoid the

continuous and insidious $\operatorname{erosion}^{43}$ of forests by fire throughout the tropics, an increased level of attention, study and resources need to be directed to tropical fire science. This cannot be done by a blanket application of current fire knowledge from other parts of the world to the tropics, but neither does fire science need to be reinvented anew in tropical guise. Instead, current fire knowledge needs to be interpreted in the context of tropical forests and, where necessary, added to, by defining the mechanisms by which fire and ecosystem processes interact⁷⁷ in these forests.

doi:10.1038/nature01437.

- Barber, C. V. & Schweithelm, J. Trial by Fire: Forest Fire and Forestry Policy in Indonesia's Era of Crisis and Reform (World Resources Institute, Washington DC, 2000).
- Cochrane, M. A. Spreading Like Wildfire—Tropical Forest Fires in Latin America and the Caribbean: Prevention, Assessment and Early Warning (United Nations Environment Programme (UNEP), Mexico City, 2002); available at http://www.rolac.unep.mx/dewalac/eng/fire_ingles.pdf.
- Ruitenbeek, J. in Indonesia's Fire and Haze: the Cost of Catastrophe (eds Glover, D. & Jessup, T.) 88–112 (International Development Research Centre, Ottawa, 1999).
- United Nations Environment Programme GEO Latin America and the Caribbean: Environment Outlook 2000 (UNEP, Mexico City, 2000).
- Barbosa, R. I. & Fearnside, P. M. Incendios na Amazonia Brasileira: estimative da emissão de gases do efeito estufa pela queima de diferentes ecossistemas de Roraima na passagem do evento El Niño (1997/ 98). Acta Amazon. 29, 513–534 (1999).
- Goldammer, J. G. et al. in Fire in the Tropical Biota (ed. Goldammer, J. G.) 487–489 (Springer, Berlin, 1990).
- Cochrane, M. A. & Schulze, M. D. Fire as a recurrent event in tropical forests of the eastern Amazon: effects on forest structure, biomass, and species composition. *Biotropica* 31, 2–16 (1999).
- Gerwing, J. J. Degradation of forests through logging and fire in the eastern Brazilian Amazon. For. Ecol. Mgmt 157, 131–141 (2002).
- Slik, J. W., Verburg, R. W. & Kebler, P. J. A. Effects of fire and selective logging on the tree species composition of lowland dipterocarp forest in East Kalimantan, Indonesia. *Biodiv. Conserv.* 11, 85–98 (2002).
- Van Nieuwstadt, M. G. L., Sheil, D. & Kartawinata, K. The ecological consequences of logging in the burned forests of East Kalimantan, Indonesia. *Conserv. Biol.* 15, 1183–1186 (2001).
- Kinnaird, M. F. & O'Brien, T. G. Ecological effects of wildfire on lowland rainforest in Sumatra. Conserv. Biol. 12, 954–956 (1998).
- 12. Kellman, M. & Meave, J. Fire in the tropical gallery forests of Belize. J. Biogeogr. 24, 23–24 (1997).
- Cochrane, M. A. et al. Positive feedbacks in the fire dynamic of closed canopy tropical forests. Science 284, 1832–1835 (1999).
- 14. Swaine, M. D. Characteristics of dry forest in West Africa and the influence of fire. J. Veg. Sci. 3, 365–374 (1992).
- Woods, P. Effects of logging, drought, and fire on structure and composition of tropical forests in Sabah, Malaysia. *Biotropica* 21, 290–298 (1989).
- Fredericksen, N. J. & Fredericksen, T. S. Terrestrial wildlife responses to logging and fire in a Bolivian tropical humid forest. *Biodiv. Conserv.* 11, 27–38 (2002).
- Barlow, J., Haugaasen, T. & Peres, C. A. Effects of ground fires on understorey bird assemblages in Amazonian forests. *Biol. Conserv.* 105, 157–169 (2002).
- Page, S. S., Siegert, F., Rieley, J. O., Boehm, H. V. & Jaya, A. The amount of carbon released from peat and forest fires in Indonesia during 1997. *Nature* 420, 61–65 (2003).
- Cairns, M. A., Hao, W. M., Alvarado, E. & Haggerty, P. in Vol. 1: Proc. Joint Fire Science Conf. and Workshop: Crossing the Millennium: Integrating Spatial Technologies and Ecological Principles for a New Age in Fire Management (eds Neuenschwander, L. F., Ryan, K. C., Gollberg, G. E. & Greer, J. D.) 242–247 (University of Idaho and the International Association of Wildland Fire, Moscow, Idaho, 2000).
- Phulpin, T., Lavenu, F., Bellan, M. F., Mougenot, B. & Blasco, F. Using SPOT-4 HRVIR and VEGETATION sensors to assess impact of tropical forest fires in Roraima, Brazil. *Int. J. Remote Sens.* 23, 1943–1966 (2002).
- Houghton, J. T. (ed.) Climate Change 2001: The Scientific Basis. Contribution of Working Group I to the Third Assessment Report of the Intergovernmental Panel on Climate Change (Cambridge Univ. Press, Cambridge, UK, 2001).
- Kunii, O. in *Health Guidelines for Vegetation Fire Events: Background Papers* (eds Goh, K., Schwela, D., Goldammer, J. G. & Simpson, O.) 299–316 (World Health Organization, Geneva, 1999).
- Linden, E. A Estrada do desastre. Time Latina (20 September 2000); available at http://cnnbrasil.com/ 2000/time/09/20/amazon/.
- Ward, D. E. in *Health Guidelines for Vegetation Fire Events: Background Papers* (eds Goh, K., Schwela, D., Goldammer, J. G. & Simpson, O.) 71–86 (World Health Organization, Geneva, 1999).
- Crutzen, P. J. & Andreae, M. O. Biomass burning in the tropics: impact on atmospheric chemistry and biogeochemical cycles. *Science* 250, 1669–1677 (1990).
- Ramanathan, V., Crutzen, P. J., Kiehl, J. T. & Rosenfeld, D. Aerosols, climate and the hydrological cycle. Science 294, 2119–2124 (2001).
- Kaufman, Y. J., Tanré, D. & Boucher, O. A satellite view of aerosols in the climate system. *Nature* 419, 215–223 (2002).
- Rosenfeld, D. TRMM observed first direct evidence of smoke from forest fires inhibiting rainfall. *Geophys. Res. Lett.* 26, 3105–3108 (1999).
- 29. Ackerman, A. S. et al. Reduction of tropical cloudiness by soot. Science 288, 1024–1047 (2000).
- Pyne, S. J. World Fire: The Culture of Fire on Earth (Univ. Washington Press, Seattle, Washington, 1997).
- Meggers, B. J. Archeological evidence for the impact of Mega-Niño events on Amazonia during the past two millennia. *Clim. Change* 28, 321–338 (1994).
- Nepstad, D. C., Moreira, A. G. & Alencar, A. A. Flames in the Rain Forest: Origins, Impacts, and Alternatives to Amazonian Fires (World Bank, Brasilia, Brazil, 1999).

NATURE | VOL 421 | 27 FEBRUARY 2003 | www.nature.com/nature

33. Tate, G. H. H. Life zones at Mount Roraima. Ecology 13, 235-257 (1932).

- Goldammer, J. G. & Seibert, B. in *Fire in the Tropical Biota* (ed. Goldammer, J. G.) 11–31 (Springer, Berlin, 1990).
- Uhl, C., Kauffman, J. B. & Cummings, D. L. Fire in the Venezuelan Amazon 2: Environmental conditions necessary for forest fires in the evergreen rainforest of Venezuela. *Oikos* 53, 176–184 (1988).
- Sanford, R. L., Saldarriaga, J., Clark, K., Uhl, C. & Herrera, R. Amazon rainforest fires. Science 227, 53–55 (1985).
- Goldammer, J. G. & Seibert, B. Natural rain-forest fires in Eastern Borneo during Pleistocene and Holocene. Naturvissenschaften 76, 518–520 (1989).
- Hammond, D. S. & ter Steege, H. Propensity for fire in Guianan rainforests. Conserv. Biol. 12, 944–947 (1998).
- Kauffman, J. B. & Uhl, C. in *Fire in the Tropical Biota* (ed. Goldammer, J. G.) 117–134 (Springer, Berlin, 1990).
- Uhl, C. & Kauffman, J. B. Deforestation, fire susceptibility, and potential tree responses to fire in the eastern Amazon. *Ecology* 71, 437–449 (1990).
- Mueller-Dombois, M. Proc. Conf. Fire Regimes and Ecosystem Properties. General Technical Report WO-26, 137–176 (US Forest Service, Washington DC, 1981).
- Moreira, M. Z. *et al.* Contribution of transpiration to forest ambient vapour based on isotopic measurements. *Glob. Change Biol.* 3, 439–450 (1997).
- 43. Aubréville, A. M. A. The disappearance of the tropical forests of Africa. Unasylva 1, 5-11 (1947).
- 44. Eltahir, E. A. B. & Bras, R. L. Precipitation recycling. *Rev. Geophys.* 34, 367–378 (1996).
- Nepstad, D. C. *et al.* The role of deep roots in water and carbon cycles of Amazonian forests and pastures. *Nature* 372, 666–669 (1994).
- Kleidon, A. & Lorenz, S. Deep roots sustain Amazonian rainforest in climate model simulations of the last ice age. *Geophys. Res. Lett.* 28, 2425–2428 (2001).
- Uhl, C. & Buschbacher, R. A disturbing synergism between cattle ranch burning practices and selective tree harvesting in the Eastern Amazon. *Biotropica* 17, 265–268 (1985).
- Cochrane, M. A. Synergistic interactions between habitat fragmentation and fire in evergreen tropical forests. *Conserv. Biol.* 15, 1515–1521 (2001).
- Cochrane, M. A. & Laurance, W. F. Fire as a large-scale edge effect in Amazonian forests. J. Trop. Ecol. 18, 311–325 (2002).
- Cochrane, M. A., Skole, D. L., Matricardi, E. A. T., Barber, C., Chomentowski, W. in Working Forests in the Tropics: Conservation through Sustainable Management? (eds Zarin, D. J. et al.) (Columbia Univ. Press, New York, in the press).
- Bucini, G. & Lambin, E. F. Fire impacts on vegetation in Central Africa: a remote-sensing-based statistical analysis. *Appl. Geogr.* 22, 27–48 (2002).
- 52. Stott, P. Combustion in tropical biomass fires: a critical review. Prog. Phys. Geogr. 24, 355–377 (2000).
- Tutin, C. E. G., White, L. J. T. & Mackangamissandzou, A. Lightning strike burns large forest tree in the Lopé Reserve, Gabon. *Glob. Ecol. Biogeogr. Lett.* 5, 36–41 (1996).
- Vayda, A. P. Finding Causes of the 1997–98 Indonesian Forest Fires: Problems and Possibilities (World Wide Fund for Nature (WWF), Jakarta, Indonesia, 1999).
- Skole, D. & Tucker, C. J. Tropical deforestation and habitat fragmentation in the Amazon: satellite data from 1978 to 1988. *Science* 260, 1905–1910 (1993).
- Laurance, W. F. et al. The future of the Brazilian Amazon: Development trends and deforestation. Science 291, 438–439 (2001).
- Laurance, W. F. et al. Biomass collapse in Amazonian forest fragments. Science 278, 1117–1118 (1997).
 Nascimento, H. E. M. & Laurance, W. F. Total aboveground biomass in central Amazonian rainforests:
- a landscape-scale study. *For. Ecol. Mgmt* **168**, 311–321 (2002). 59. Veríssimo, A., Barreto, P., Tarifa, R. & Uhl, C. Extraction of a high-value natural source from Amazon:
- the case of mahogany. For. Ecol. Mgmt 72, 39–60 (1995). 60. Holdsworth, A. R. & Uhl, C. Fire in Amazonian selectively logged rain forest and the potential for fire
- reduction. *Ecol. Appl.* 7, 713–725 (1997).
 61. Siegert, F., Ruecker, G., Hinrichs, A. & Hoffman, A. A. Increased damage from fires in logged forests
- during droughts caused by El Niño. Nature 414, 437-440 (2001).
- Cochrane, M. A. In the line of fire: Understanding the impacts of tropical forest fires. *Environment* 43, 28–38 (2001).
- Gascon, C., Williamson, G. B. & Fonseca, G. A. B. Receding edges and vanishing fragments. *Science* 288, 1356–1358 (2000).
- Siegert, F. & Ruecker, G. Use of multitemporal ERS-2 SAR images for identification of burned scars in south-east Asian tropical rainforest. *Int. J. Remote Sens.* 21, 831–837 (2000).
- Huber, O., Steyermark, J. A., Prance, G. T. & Alés, C. The vegetation of the Sierra Parima, Venezuela-Brazil: Some results of recent exploration. *Brittonia* 36, 104–139 (1984).
- Nepstad, D. C., Jipp, P., Moutinho, P., Negreiros, G. & Vieira, S. in Evaluating and Monitoring the Health of Large-Scale Ecosystems (eds Rapport, D., Gander, C. & Calow, P.) 333–349 (Springer, New York, 1995).
- Anderson, H. E. Aids to Determining Fuel Models for Estimating Fire Behavior. General Technical Report INT-122 (US Forest Service, Ogden, Utah, 1982).

- Kauffman, J. B., Cummings, D. L. & Ward, D. E. Fire in the Brazilian Amazon 2. Biomass, nutrient pools and losses in cattle pastures. *Oecologia* 113, 415–427 (1998).
- Finney, M. A. FARSITE: Fire Area Simulator—Model Development and Evaluation (US Forest Service, Ogden, Utah, 1998).
- Gould, K. A. et al. Post-fire tree regeneration in lowland Bolivia: implications for fire management. For. Ecol. Mgmt 165, 225–234 (2002).
- Gerwing, J. J. Testing liana cutting and controlled burning as silvicultural treatments for a logged forest in the eastern Amazon. J. Appl. Ecol. 38, 1264–1276 (2001).
- 72. Quintere, J. G. Canadian mass fire experiment. 1989. J. Fire Protect. Eng. 5, 67-78 (1993).
- Bailey, W. O. Report on the Michigan Forest Fires of 1881 (US War Department, Washington DC, 1882).
 Laurance, W. F. & Williamson, B. Positive feedbacks among forest fragmentation, drought, and climate change in the Amazon. Conserv. Biol. 15, 1529–1535 (2001).
- Lawton, R. O., Nair, U. S., Pielke, R. A. & Welch, R. M. Climatic impact of tropical lowland deforestation on nearby montane cloud forests. *Science* 294, 584–587 (2001).
- Lyons, W. A., Nelson, T. E., Williams, E. R., Cramer, J. A. & Turner, T. R. Enhanced positive cloud-toground lightning in thunderstorms ingesting smoke from fires. *Science* 282, 77–80 (1998).
- 77. Johnson, E. A. & Miyanishi, K. Forest Fires: Behavior and Ecological Effects (Academic, San Diego, California, 2001).
- Veríssimo, A., Cochrane, M. A. & Souza, C. Jr National forests in the Amazon. Science 297, 1478 (2002).
- DeFries, R. S., Bounoua, L. & Collatz, G. J. Human modification of the landscape and surface climate in the next fifty years. *Glob. Change Biol.* 8, 438–458 (2002).
- Giglio, L., Kendall, J. D. & Justice, C. O. Evaluation of global fire detection algorithms using simulated AVHRR infrared data. *Int. J. Remote Sens.* 20, 1947–1985 (1999).
- Elvidge, C. D. et al. in Remote Sensing Change Detection: Environmental Monitoring Methods and Applications (eds Lunetta, R. S. & Elvidge, C. D.) 74–122 (Ann Arbor Press, Ann Arbor, Michigan, 1999).
- Fuller, D. O. & Fulk, M. Comparison of NOAA-AVHRR and DMSP-OLS for operational fire monitoring in Kalimantan, Indonesia. *Int. J. Remote Sens.* 21, 181–187 (2000).
- Pereira, A. C. & Setzer, A. W. Comparison of fire detection in savannas using AVHRR's channel 3 and TM images. *Int. J. Remote Sens.* 17, 1925–1937 (1996).
- Fuller, D. O. Satellite remote sensing of biomass burning with optical and thermal sensors. *Prog. Phys. Geogr.* 44, 543–561 (2000).
- Giglio, L., Kendall, J. D. & Tucker, C. J. Remote sensing of fires with the TRMM VIRS. *Int. J. Remote Sens.* 21, 203–207 (2000).
- Goldammer, J. G. in *Health Guidelines for Vegetation Fire Events: Background Papers* (eds Goh, K., Schwela, D., Goldammer, J. G. & Simpson, O.) 9–70 (World Health Organization, Geneva, 1999).
- Oertel, D. et al. in Forest Fire Monitoring and Mapping: a Component of Global Observation of Forest Cover (eds Ahern, F., Grégoire, J. M. & Justice, C.) 224–228 (Joint Research Centre, Ispra, Italy, 2000).
 Eva, H. & Lambin, E. Burnt area mapping in Central Africa using ATSR data. Int. J. Remote Sens. 19,
- tota, it is built area mapping in Central Array using PUSK data. Int. J. Remote Sens. 19, 3473–3497 (1998).
 Cochrane, M. A. & Souza, C. M. Jr Linear mixture model classification of burned forests in the eastern
- Amazon. Int. J. Remote Sens. 19, 3433–3440 (1998).
- Stone, T. & Lefebvre, P. Using multi-temporal satellite data to evaluate selective logging in Pará. Brazil. Int. J. Remote Sens. 19, 2517–2526 (1998).
- Nepstad, D. C. *et al.* Large-scale impoverishment of Amazonian forests by logging and fire. *Nature* 398, 505–508 (1999).
- Matricardi, E. A. T., Skole, D. L., Chomentowski, W. & Cochrane, M. A. Multi-Temporal Detection and Measurement of Selective Logging in the Brazilian Amazon Using Remote Sensing (Michigan State Univ., East Lansing, Michigan, 2001); available at http://www.globalchange.msu.edu/publications.
- Uhl, C. et al. Natural resource management in the Brazilian Amazon. Bioscience 47, 160–168 (1997).
 Jackson, S. M., Fredericksen, T. S. & Malcolm, J. R. Area disturbed and residual stand damage
- following logging in a Bolivian tropical forest. For Ecol. Mgmt 166, 271–283 (202).
- Sist, P. & Nguyen-Thé, N. Logging damage and the subsequent dynamics of a dipterocarp forest in East Kalimantan. For. Ecol. Mgmt 165, 85–103 (2002).
- Goldammer, J. G., Seibert, B. & Schindele, W. in Dipterocarp Forest Ecosystems: Towards Sustainable Management (eds Schulte, A. & Schöne, D.) 155–185 (World Scientific, Singapore, 1996).
- 97. Agee, J. K. Fire Ecology of Pacific Northwest Forests (Island Press, Washington DC, 1993).

Acknowledgements I am grateful for support from NASA.

Correspondence and requests for materials should be addressed to the author (e-mail: cochrane@globalchange.msu.edu).

review article